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Impact of population and latrines on fecal contamination of ponds in rural Bangladesh

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ABSTRACT

A majority of households in Bangladesh rely on pond water for hygiene. Exposure to pond water fecal contamination could therefore still contribute to diarrheal disease despite the installation of numerous tubewells for drinking. The objectives of this study are to determine the predominant sources (human or livestock) of fecal pollution in ponds and examine the association between local population, latrine density, latrine quality and concentrations of fecal bacteria and pathogens in pond water. Forty-three ponds were analyzed for *E. coli* using culture-based methods and *E. coli*, Bacteroidales and adenovirus using quantitative PCR. Population and sanitation spatial data were collected and measured against pond fecal contamination. Humans were the dominant source of fecal contamination in 79% of the ponds according to Bacteroidales measurements. Ponds directly receiving latrine effluent had the highest concentrations of fecal indicator bacteria (up to 10⁶ Most Probable Number (MPN) of culturable *E. coli* per 100 mL). Concentrations of fecal indicator bacteria correlated with population surveyed within a distance of 30–70 m ($p < 0.05$) and total latrines surveyed within 50–70 m ($p < 0.05$). Unsanitary latrines (visible effluent or open pits) within the pond drainage basin were also significantly correlated to fecal indicator concentrations ($p < 0.05$). Water in the vast majority of the surveyed ponds contained unsafe levels of fecal contamination attributable primarily to unsanitary latrines, and to lesser extent, to sanitary latrines and cattle. Since the majority of fecal pollution is derived from human waste, continued use of pond water could help explain the persistence of diarrheal disease in rural South Asia.

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1. Introduction

Over the past few decades the incidence of diarrheal disease has declined throughout the developing world. Despite this decline, 1.5 million children under five die of diarrhea annually (UNICEF/WHO, 2009). The decline in diarrheal disease has generally been attributed to improved medical treatment and interventions focused on safe drinking water, as well as improved sanitation and hygiene (Esrey, 1996; Pruss et al., 2002). The inability to reduce diarrheal disease levels

further suggests that transmission pathways in the developing world are complex and that environmental conditions may play a larger role than previously thought. Substantial amounts of surface water are contained in ponds scattered throughout villages in rural Bangladesh. Due to poor sanitation, ponds receive fecal contamination from numerous latrines and could be a source of pathogens overlooked by intervention programs focused exclusively on drinking water quality. Contact with pond water has previously been identified as a driver of diarrheal disease in Bangladesh when people drink, bathe in, or live near a pond (Emch et al., 2008; Ali et al., 2002). Although surface water from ponds is generally not used for drinking and cooking, it is used for bathing and brushing teeth (Aziz et al., 1990).

In recent decades, the number of ponds excavated in Bangladesh has outpaced population growth (see Supp. Material in Neumann et al., 2010). While many ponds are excavated to raise the ground to

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protect dwellings from flooding, they often fulfill alternate purposes, including aquaculture, bathing, irrigation, or holding latrine effluent. In the area surrounding the ponds, latrines, the primary mode of handling human feces in rural Bangladesh, are often deliberately placed, effectively creating sewage lagoons. Global studies have revealed that substantial decreases in diarrheal disease morbidity by a quarter to over a third (Fewtrell et al., 2005) and improvements in childhood nutritional indexes (Esrey, 1996) have accompanied a gradual switch from open pit latrines to more sanitary latrines consisting of concrete foundation rings to prevent leakage of human feces onto the open ground. In rural Bangladesh specifically, epidemiologic studies have reported reductions in diarrheal disease morbidity due to improved sanitation (Aziz et al., 1990; Emch, 1999; Emch et al., 2008; Hoque et al., 1996) where only half the population currently has access to sanitary latrines (WHO/UNICEF, 2008).

Although several diarrheal disease pathogens are transmitted through human feces including, *Shigella* and rotavirus, livestock feces can carry diarrheal pathogens including *Campylobacter*, *Salmonella* and certain types of pathogenic *E. coli* (Nicholson et al., 2005), all of which are routinely identified in the stools of diarrhea patients at the International Center for Diarrheal Disease Research Bangladesh (ICDDR) hospital (Albert et al., 1999). Villagers live very close to their livestock on a daily basis and the livestock are frequently kept near ponds where latrines and tubewells are also located, resulting in opportunities for both types of fecal contamination to impact pond water quality. Therefore, understanding the relative contributions of fecal sources, as well as the impact of waste containment on surface water fecal contamination, could lead to better public health interventions.

The goals of this study were to assess the levels and factors influencing concentrations of fecal bacteria and adenovirus in ponds within a representative village of Bangladesh. The following hypotheses were tested: 1) fecal bacteria in ponds of rural Bangladesh are primarily human in origin, rather than from domestic livestock; 2) concentrations of fecal bacteria and adenovirus are related to local population density; and 3) concentrations of fecal bacteria and adenovirus are related to the number of latrines around a pond, as well as the type of latrine. To test these hypotheses, this study integrates pond concentrations of fecal bacteria and viral pathogens with detailed spatial data from the same village. The novelty of this study is: 1) both *E. coli* and total Bacteroidales were compared as fecal indicators; 2) host animal specific Bacteroidales quantitative PCR assays were used to identify samples impacted by human or livestock fecal waste (Bernhard and Field, 2000; Layton et al., 2006; Lee et al., 2008); and 3) human adenovirus, representing both a human pathogen and a human fecal indicator, was analyzed (Castignolles et al., 1998; Wyn-Jones and Sellwood, 2001; Jiang, 2006).

2. Methods

2.1. Site description

The village of Char Para is located in Arahazar upazilla, about 25 km east of Dhaka. On-going public-health and earth-science studies focused on the groundwater arsenic problem were launched in the area in 2000 (van Geen et al., 2003). Char Para, also referred to as Site K (Radloff et al., 2007), is underlain by fine to medium grained deltaic sands, which form a shallow aquifer that is tapped by tubewells that are screened from 10 to 20 m and are the primary drinking water source in the village. The shallow aquifer below Char Para is bounded hydrologically on three sides by a former channel of the Old Brahmaputra River which floods to the edge of the village by the late wet season (van Geen et al., 2003; Weinman et al., 2008). Many ponds in Char Para are empty at the end of the dry season in April when the groundwater table falls below the bottom of the pond. The ponds that do not dry out are often the deepest ponds, or are

artificially maintained for fish farming by pumping from a deeper aquifer. Standing water can also be maintained year-round in latrine ponds that receive latrine effluent and runoff of wash water from surrounding wells. At the beginning of the monsoon in late May, some ponds may also fill and drain rapidly; fluctuations in pond water levels of up to 1 m were observed within 24 h in June 2008. As a result of the initially depressed water table and the flat topography between pond basins, little mixing between surface water bodies occurs until late in the monsoon (August) when some ponds may overflow their basins.

2.2. Field methods and pond classification

2.2.1. Village-wide GPS survey

High accuracy (sub-meter) GPS coordinates were collected for all ponds, latrines and households throughout the village during June 2009 using a Trimble GeoXH receiver and Terrasync 2.4 software (Table A1). GPS data were post-processed using Pathfinder Office 3.0 (Trimble Navigation Ltd., Sunnyvale, CA). Latrines were classified as sanitary if the concrete platform and rings were intact with no visible sign of effluent discharging onto the ground. A latrine was classified as unsanitary if the ring was absent or cracked, or if visible effluent discharged directly into a pond via a PVC pipe. A population survey was conducted to determine pond ownership and the number of persons living in each household (cattle were not counted during the village-wide survey). Population and latrine totals within a given radius of each pond were determined using the buffer and intersect tools in ArcGIS software. The population and number of latrines (sanitary and unsanitary) surrounding ponds within a radius of 10 and 50 m was calculated at 5-m intervals and within a radius of 50 to 100 m at 10-m intervals. This method is referred to hereon as the spatial buffer method.

2.2.2. Pond drainage basin survey

An alternative survey was carried out in June 2008 to identify latrines within a radius of ~20 m of a pond located within drainage basin that sloped downward toward the water edge. The rationale for this hydrological approach is that latrines within the basin could have a greater influence on microbial pond water quality than latrines at similar distances outside the drainage basin. Information collected for each pond drainage basin includes water depth, long and short axes of the pond water surface (measured in meters), designated purpose as identified by the owner, number and type of latrines (unsanitary or sanitary) and the number of cattle residing within the drainage basin (Table A1). Ponds receiving direct latrine effluent were classified as latrine ponds. Unless local households identified a specific use such as bathing or aquaculture, ponds not receiving direct latrine effluent were categorized as having no specific use. Specific conductivity, pH, dissolved oxygen and temperature were measured at each pond at the time of microbial sampling using a calibrated handheld multiprobe (556 Multiprobe System, YSI Inc., Yellow Springs, OH).

2.3. Microbiological assays

Water from 43 ponds was collected in sterile bottles in mid-June 2008, during the early monsoon when surface runoff was common but the ponds were not yet full. One hundred milliliters of pond water samples were collected in sterile containers to measure culturable *E. coli* using the MPN based Colilert™ test kit (IDEXX Laboratories, Inc., Westbrook, ME). Triplicate samples were taken from the first 30 ponds, and after the Coefficient of Variation (C.V.) between replicates was found to be low (24%), remaining ponds were sampled in duplicate (C.V. 31%). Based on the assumption that replicates contained the same true concentration of cultured *E. coli* cells, triplicate and duplicate samples were pooled to produce a single MPN with 95% confidence intervals (Hurley and Roscoe, 1983; Knappett et al., 2010). Pond water samples were diluted 1:100 or 1:1000 with commercial bottled drinking water

before being assayed to avoid exceeding the quantifiable maximum of 2419 bacteria/100 mL of the assay. The minimum detection limit of a single Colilert™ analysis is 1 MPN/100 mL, however since each sample was diluted at least 100 times, the detection limit became 100 MPN/100 mL. All but 1 pond had culturable *E. coli* concentrations above the 100 MPN/100 mL detection limit. Laboratory blanks using bottled water were included every thirty 100 mL samples to control for contamination during processing.

For the molecular measurements as much pond water as could be filtered before clogging, was filtered through a 0.22 µm nitrocellulose filter (150 mL Vacuum Driven Disposable Filtration System, Stericup, HV Durapore Membrane, Millipore Corp., Bedford, MA). The volumes filtered ranged from 40 to 295 mL. Filters were removed from the plastic housing, placed in sterile petri plates, kept on ice less than 8 h and then frozen for ~1 week at –10 °C and transported on dry ice to the University of Tennessee where they were stored at –80 °C. DNA was extracted and purified from the filters using the FastDNA® SPIN for Soil Kit (MP Biomedicals, LLC, Solon, OH) following the manufacturer's protocols. The original extracted DNA was quantified using a NanoDrop Spectrophotometer (ThermoScientific, Wilmington, DE) and diluted up to 40 fold prior to qPCR to obtain a concentration between 5 and 10 ng/µL in order to minimize potential PCR inhibition.

Quantitative PCR (qPCR) was performed using assays designed for *E. coli* and total Bacteroidales, three host-associated Bacteroidales assays (two for human and one for bovine), and the human pathogen adenovirus (Table 2). The gene targets for the *E. coli* and Bacteroidales assays were the 23S rRNA gene (Smith et al., 1999; Layton et al., 2003) and the 16S rRNA gene, respectively, with the human and bovine host-associated assays targeting different subgroups within the Bacteroidales order (Bernhard and Field, 2000; Layton et al., 2006; Seurinck et al., 2005). Genome equivalents of *E. coli* (using the 23S rRNA molecular target) detected in a water sample are hereon referred to as *mE. coli* (molecular) to distinguish them from cultured *E. coli*. The adenovirus gene target was the hexon gene from serotypes 40 and 41 (Rajal et al., 2007) which encodes the major capsid protein for the adenovirus (Ebner et al., 2005).

All qPCR reactions consisted of 12.5 µL PCR mix (Stratagene, LaJolla, CA), 0.75 µL of a 20 µM stock of each of the forward primer and reverse primers, and 0.5 µL of a 10 µM stock of the probe (Table 2), 8 µL of sterile water and 2.5 µL of sample or standard for a final reaction volume of 25 µL. PCR amplification and fluorescent probe detection were performed using the Chromo4 Real-Time PCR Detection system (BioRad,

Hercules, CA) and the following amplification protocol: 50 °C for 2 min, 95 °C for 10 min, and 45 cycles of alternating 95 °C for 30 s and the annealing temperatures as indicated in Table 1 for 45 s. Calibration of the qPCR assays using a standard curve involved cloning the target gene into a plasmid for all assays except for the *mE. coli* assays which used *E. coli* O157 genomic DNA (Strain EDL933, ATCC 43895D-5). For plasmid standards, serial 10× dilutions were performed in triplicate from a starting concentration of 1 × 10⁷ plasmid copies to 10 copies per µL and 2.5 µL of each plasmid dilution were placed in triplicate wells. Similarly the *E. coli* O157 genomic DNA was diluted serially from a starting concentration of 1 × 10⁶ to 10 genomic equivalents per mL and 2.5 µL of each plasmid dilution was placed in triplicate wells. All qPCR assays were performed in triplicate for each sample with a fourth well spiked with 2.5 × 10⁵ copies of the standard in order to monitor PCR inhibition using Eq. (A.1) to calculate recovery efficiency as described previously (Bell et al., 2009; Layton et al., 2006). Recovery efficiency is also referred to as the inhibition factor by others (Yamahara et al., 2009). One sample had PCR inhibition, so it was removed from all subsequent data analysis. In rare cases where the standard deviation exceeded the mean gene concentration in the 3 wells (C.V. > 100%), the assay for that marker was re-run. Final concentrations of target copies in original water samples were calculated using Eq. (A.2).

The method detection limit (MDL in copies/100 mL) was variable due to dilution of extracted DNA to limit inhibition that may accompany water samples with high biomass. The MDL for each sample was determined when the copies of target DNA were less than 1 copy per ng of total DNA in the qPCR reaction well. Each extracted DNA sample was diluted down to 5 to 10 ng/µL and therefore the limit for each sample (2.5 µL) ranged from 12.5 to 25 copies per qPCR reaction, which was within the range of quantification on the standard curve. Since the initial concentration of total DNA extracted from the samples varied from 8 to 413 ng/µL and the volume of pond water filtered ranged from 40 to 295 mL, the MDL also varied between samples (Eq. (A.2)) with a geometric mean of 8.4 × 10³ and a range of 1.4 × 10³ to 2.1 × 10⁵ copies/100 mL. For each sample, this MDL was the same for all molecular targets.

2.4. Statistical analysis

Concentrations of all six markers (culturable *E. coli*, *mE. coli*, total Bacteroidales, human and bovine Bacteroidales and adenovirus) were compared using a non-parametric correlation matrix which reports the

Table 1

Real-time PCR assays used to detect *E. coli*, and Bacteroidales rRNA genes and the adenovirus hexon gene, the primers and probe used for each assay, and the annealing temperature used for each assay.

Assay name (target organism)	Primer/probe name and sequence (5'–3')	Size (bp) of product	Annealing temp (°C)	Reference
EC23S (<i>E. coli</i>)	EC23Sf, 5'-GAG CCT GAA TCA GTG TGT GTG-3' EC23Sr, 5'-ATT TTT GTG TAC GGG GCT GT-3' EC23Srv1bhq, 5'-(FAM)CGC CTT TCC AGA CGC TTC CAC(BHQ-1)-3'	78	55	Modified from Smith et al. (1999), Layton et al. (2003)
AllBac (all Bacteroidales)	AllBac296f, 5'-GAGAGGAAGGTCCCCAC-3' AllBac412r, 5'-CGCTACTTGGCTGGTTCAG-3' AllBac375Bhqr, 5'-(FAM)CCATTGACCAATATTCCTCACTGCTGCCT(BHQ-1)-3'	106	60	Layton et al. (2006)
HuBac	mHuBac563f, 5'-ATTGGGTTAAAGGGAGCGTAG-3' mHuBac694r, 5'-CTACACCACGAATCCGCC-3' mHuBac594Taq, 5'-(FAM)TAAGTCAGTTGTGAAAGTTTGGCGCTC(BHQ-1)-3'	131	69	Layton et al. (2006)
HF183-like	GBAC34f 5'-CGC TAG CTA CAG GCT TAA CAC-3' GBAC313r, 5'-GTG GGG GAC CTT CCT CTC-3' SerH285bhq, 5'-(FAM)ATCCATCGTTGACTAGGTGGCCGTTA(BHQ-1)-3'	279	60	Modified from Bernhard and Field (2000)
BoBac	CBac367f, 5'-GAAG(G/A)CTGAACACGCAAGTA-3' CBac467r, 5'-GCTTATTCATACGGTACATACAAG-3' CBac402 Bhq, 5'-(FAM)TGAAGGATGAAGGTTCTATGGATTGAAACTT(BHQ-1)-3'	100	57	Layton et al. (2006)
Adeno (adenovirus 40/41 hexon gene)	AV40/41-117f, 5'-CAGCCTGGGAACAAGTTCAG-3' AV40/41-258r, 5'-CAGCGTAAAGCGCACTTTGTAA-3' AV40/41-157BHQ, 5'-(FAM)ACCCACGATGTAACACAGACAGGTC(BHQ-1)-3'	141	60	Rajal et al. (2007)

Spearman Rank Order Correlation Coefficient r_s and significance level ($p < 0.05$). The non-parametric Kruskal–Wallis test was performed using NCSS software (version 07.1.14, NCSS, LLC, Kaysville, UT) to determine concentration differences between pond groups based on type ($p < 0.05$). Associations between population and spatial-buffer latrine counts (sanitary, unsanitary, and total) within 10 to 100 m of a pond and concentrations of pond water markers were also tested using the Spearman Rank Order Coefficient. For samples Below Detection Limit (BDL) the geometric mean qPCR detection limit (8.4×10^3 copies/100 mL) was substituted. Half (11) of the total BDL's (23) were for Bovine Bacteroidales (26% substitution rate) indicating that correlation statistics reported for Bovine Bacteroidales here should be considered as semi-quantitative estimates (Table A2). For the other markers this method of data censoring likely had little impact on the reported statistics with only 5% (12/215) of marker analyses BDL (Helsel, 2006; Antweiler and Taylor, 2008).

3. Results

3.1. Physical and chemical attributes of ponds

All 43 ponds found within the 0.3 km² area of the village were sampled in June 2008. Of the ponds sampled, 11 were designated by the owner as fish or bathing ponds and 16 had no designated purpose (Fig. 1). The remaining 16 ponds were classified as latrine ponds because they received direct effluent from at least one latrine. Pond surface area ranged from 20 to 1200 m². The largest ponds were commercial fishing and community bathing ponds, located in the northeast section of the village and contain water year round by pumping from wells (Fig. 1).

Specific conductivity may be a proxy for fecal contamination in these ponds due to high salt concentrations in animal feces (Manahan, 2005). The specific conductivity of water in latrine ponds (median 0.41 mS/cm) was found to be higher than in fish/bathing ponds (0.17 mS/cm) (Fig. 2a). Similarly, dissolved oxygen concentrations are depleted in waters receiving human or animal excrement, and in the

present study latrine ponds (median 0.24 ppm) were lower compared with fish/bathing ponds (0.68 ppm) ($p < 0.05$), while the fish/bathing ponds had the largest range of dissolved oxygen (0.23–1.51 ppm) (Fig. 2b). Temperature, a factor known to influence the persistence of fecal bacteria and their molecular markers in surface water (Bell et al., 2009), ranged from 27 to 34 °C (Fig. 2c), with higher temperature ponds located on the edges of the village or in fields, where there is little to no shade. The smaller and cooler ponds, often latrine ponds, were located in the interior of the village and typically surrounded by trees. Median specific conductivity (Fig. 2a), dissolved oxygen (Fig. 2b) and temperature (Fig. 2c), for ponds without a defined use were intermediate between median values for fish/bathing and latrine ponds.

3.2. Concentrations of indicator bacteria and genetic markers

In the 43 ponds sampled, cultured *E. coli* were detected in 42, *mE. coli* in 39, Bacteroidales in 43 and adenovirus in 41. Human and bovine Bacteroidales were detected in 36 and 24 ponds respectively (Table A2). In the ponds with detectable human and bovine Bacteroidales, copy numbers ranged from 1.3×10^4 to 6.8×10^7 and 3.3×10^4 to 7.0×10^6 copies/100 mL, respectively. Adenovirus copy numbers ranged from 1.3×10^4 to 8.8×10^6 copies/100 mL.

Most Probable Numbers of cultured *E. coli* in the 43 ponds ranged from non-detect (<100 MPN/100 mL) to 9.7×10^5 MPN/100 mL (Fig. 2d). Ponds with the highest Most Probable Numbers of culturable *E. coli* tended to be located in the central part of the village (Fig. 1). *mE. coli* genome numbers ranged from non-detect (< 8.4×10^3 copies/100 mL) to 3.2×10^7 copies/100 mL (Fig. 2e). The log₁₀-transformed concentrations of culturable *E. coli* cells and *mE. coli* genomes were fit to an exponential curve using linear regression ($R^2 = 0.43$) (Fig. 3a). The median proportion of *E. coli* culturable was 2.4% (25th, 75th percentile = 0.7, 6.1% respectively). Although pond temperature varied from 27 °C within the village to 34 °C outside, the ratio of cultured to molecular *E. coli* was uncorrelated with temperature. The correlation between *mE. coli* genomes and cultured *E. coli*, and the higher abundance of *mE. coli*

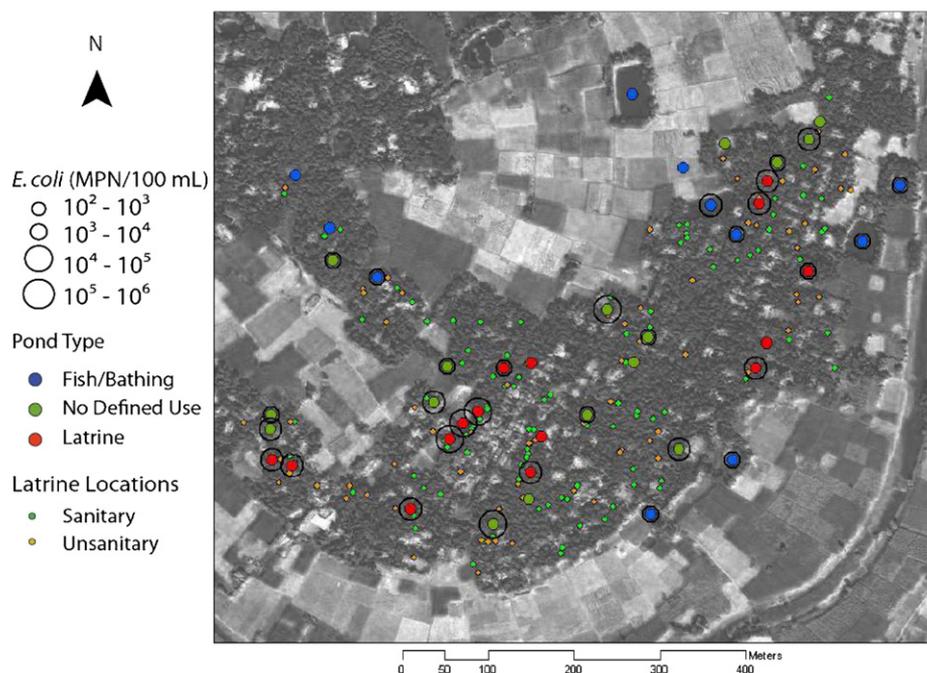


Fig. 1. Concentration of cultured *E. coli* in each pond classified by pond type with locations of sanitary and unsanitary latrines. IKONOS satellite image taken of the entire region of Arahazar at 1 m resolution (van Geen et al., 2003).

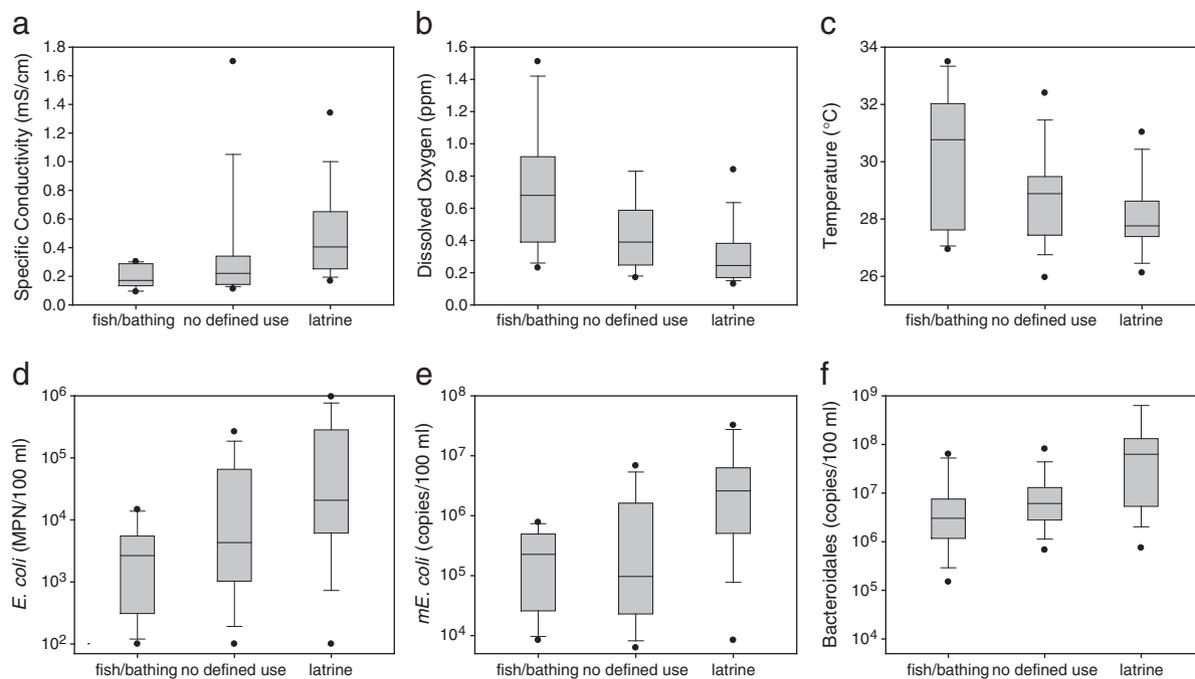


Fig. 2. Dissolved oxygen, specific conductivity, temperature and concentrations (MPN or copies/100 mL) of three fecal markers in water, from three types of ponds at Site K. The center line represents the median, upper and lower bounds of the box are the 75th and 25th percentile and whiskers represent the extent of the data. Outliers are represented by dots. The number of ponds were 43, consisting of 11 fish/bathing, 16 latrine and 16 ponds with no defined use in each category. The geometric mean detection limit for molecular assays was 8374 copies/100 mL and is indicated by the dotted line.

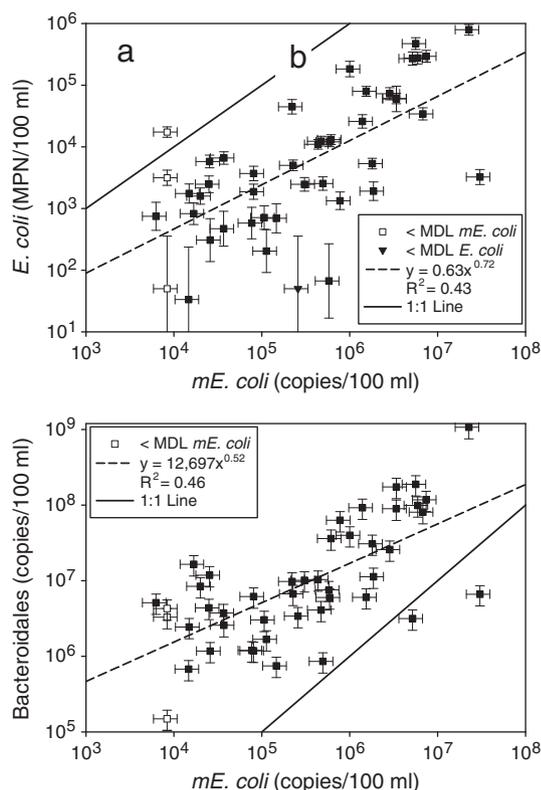


Fig. 3. Observed and predicted *E. coli* and total Bacteroidales as a function of *mE. coli* in 43 pond water samples. One non-detect occurred with culturable *E. coli* and 3 non-detects occurred for *mE. coli*, but not in the same samples. Total Bacteroidales was detected in every pond water sample. Error bars represent 95% analytical confidence intervals. Predictive equation is the result of fitting a linear regression model ($y = mx + b$) to \log_{10} -transformed concentrations with R^2 . The 1:1 line is shown for comparison ($y = x$).

genomes over cultured *E. coli* cells, supports its reliability as a quantitative indicator of fecal pollution.

As a distinct species of bacteria abundant in the intestines of warm blooded organisms Bacteroidales is a fecal indicator independent from *E. coli*. Pond water total Bacteroidales copy numbers were higher and correlated with *mE. coli* genomes in all but two cases (Fig. 3b, Table 2). In 34 out of 38 samples where at least one host-specific marker was detected, the concentration of human Bacteroidales exceeded that of bovine Bacteroidales (Fig. 4). Although total Bacteroidales were detected in all ponds, neither human nor bovine Bacteroidales were detected in 5 ponds (Fig. 4). Copy numbers of cultured *E. coli*, *mE. coli*, and total Bacteroidales in pond water were all significantly correlated (Table 2) reinforcing total Bacteroidales as a quantitative indicator of the degree of fecal contamination in the ponds. Adenovirus was present in the majority of ponds but its copy numbers were uncorrelated to fecal indicator bacteria concentrations (Table 2) and pond type (Fig. A3).

Fish/bathing ponds were significantly less contaminated than latrine ponds according to measurements on the three quantitative fecal indicator bacteria (Fig. 2d, e, and f) ($p < 0.05$). The median concentrations of fecal indicators in ponds without a defined use were intermediate of corresponding median concentrations in latrine and fish/bathing ponds. The median cultured *E. coli* MPN's (with associated 25th and 75th percentiles) for latrine, no defined use and fish/bathing ponds were 2.1×10^4 (7.9×10^3 , 2.2×10^5), 4.3×10^3 (2.1×10^3 , 4.9×10^4) and 2.7×10^3 , (5.1×10^2 , 4.4×10^3) MPN/100 mL respectively.

3.3. Population density and pond contamination

Approximately 1500 people live within Char Para. The Spearman rank correlation coefficient between the spatial-buffer population count and the concentration of the various microbial markers was calculated for distances from a pond ranging from 10 to 100 m (Fig. A2a). The association between population and fecal bacteria concentrations was maximized at a 45 m spatial-buffer radius with all fecal indicator

Table 2
Spearman Rank Order correlations of microbial markers with extracted and measured field parameters. Significance in association $p < 0.05$ is indicated by bold.

		GIS				Pond basin				Measured on pond water							
		Population (45 m)	Unsanitary latrines (60 m)	Sanitary latrines (60 m)	Total latrines (60 m)	Unsanitary latrines	Sanitary latrines	Total latrines	Cattle	Temperature (°C)	Conductivity (mS/cm)	Dissolved oxygen (mg/L)	<i>E. coli</i>	<i>mE. coli</i>	Bacteroides	Human Bacteroides	Bovine Bacteroides
GIS	Unsanitary latrines (60 m)	0.71															
	Sanitary latrines (60 m)	0.52	0.27														
	Total latrines (60 m)	0.77	0.90	0.60													
Pond basin	Unsanitary latrines	0.37	0.32	0.16	0.35												
	Sanitary latrines	0.03	0.08	0.08	0.08	0.43											
	Total latrines	0.28	0.26	0.11	0.27	0.92	0.72										
	Cattle	0.20	0.10	0.18	0.21	0.45	0.17	0.39									
Measured on pond water	Temperature (°C)	−0.30	−0.23	−0.25	−0.25	−0.21	−0.06	−0.19	−0.07								
	Conductivity (mS/cm)	0.38	0.10	0.20	0.23	0.43	0.05	0.35	0.27	−0.26							
	Dissolved oxygen (mg/L)	−0.36	−0.42	−0.10	−0.35	−0.31	−0.20	−0.33	0.11	0.69	−0.29						
	<i>E. coli</i>	0.57	0.33	0.31	0.40	0.46	0.00	0.33	0.32	−0.18	0.47	−0.14					
	<i>mE. coli</i>	0.46	0.30	0.16	0.37	0.37	−0.25	0.18	0.40	−0.04	0.49	−0.08	0.69				
	Bacteroides	0.38	0.26	0.28	0.36	0.41	0.01	0.31	0.41	−0.09	0.52	−0.11	0.69	0.69			
	Human Bacteroides	0.34	0.22	0.22	0.24	0.17	−0.04	0.09	0.05	−0.23	0.41	−0.04	0.69	0.42	0.50		
	Bovine Bacteroides	0.36	0.11	0.27	0.21	0.19	−0.05	0.10	0.05	−0.06	0.41	−0.04	0.70	0.53	0.47	0.63	
	Adenovirus	0.06	0.07	−0.18	−0.04	0.10	−0.20	0.03	0.02	0.11	0.16	0.11	0.09	0.25	0.25	0.22	0.12

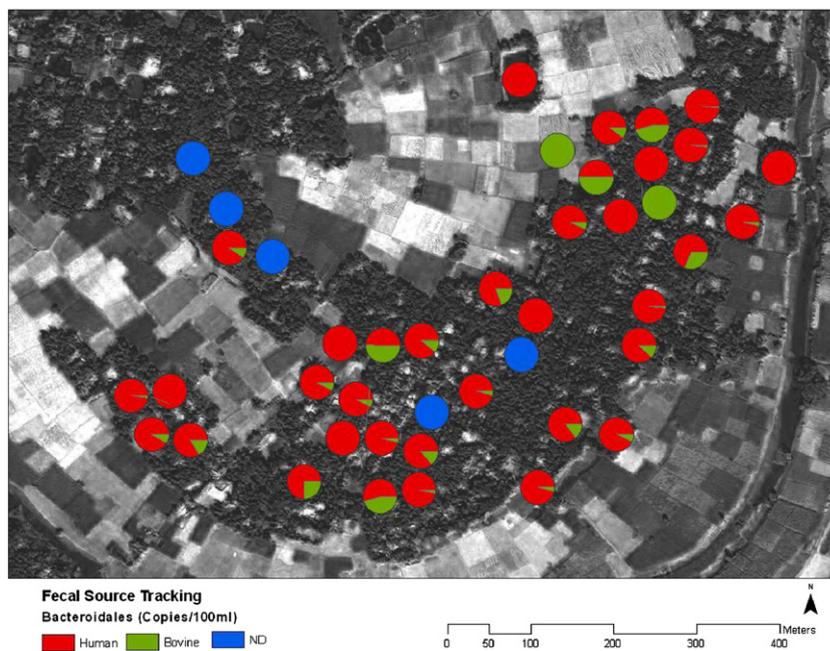


Fig. 4. Relative amounts of fecal contamination from human or bovine sources. Blue circles represent ponds where human or bovine Bacteroidales were not detected. (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article.)

bacteria showing significant correlations to population ($p < 0.05$). Population within 45 m of each pond ranged from zero to 126 people (Table A4).

High Spearman correlations are obtained for population within 45 m of each pond and all fecal markers with r_s values of 0.57, 0.46, and 0.38 for cultured *E. coli*, *mE. coli* and total Bacteroidales respectively (Table 2). Human Bacteroidales molecular targets (HuBac and HF 183), at least one of which were detected in 36 out of 43 samples, correlated significantly to the population within 45 m of the pond ($r_s = 0.34$) as did Bovine Bacteroidales ($r_s = 0.36$). In contrast, adenovirus was uncorrelated to population at any spatial-buffer distance.

3.4. Latrine density and pond contamination

A total of 178 latrines could be located during the village-wide GPS survey (Fig. 1), 79 (42%) of which were sanitary latrines, similar to the proportion observed elsewhere in Bangladesh (WHO/UNICEF, 2008). Of the 99 unsanitary latrines, 22 were open pit latrines. With respect to latrines and ponds the survey was exhaustive within the boundaries of the village, however open pit latrines were difficult to find, and therefore some may have been missed. Most Probable Numbers of cultured *E. coli* were significantly correlated with the total number of latrines located within 20 to 80 m of a pond ($p < 0.05$) (Fig. A2b). The total number of latrines ranged from 0 to 8 within a 20 m spatial-buffer and 0 to 34 within 80 m. The correlation between number of unsanitary latrines and cultured *E. coli* concentration was significant across the spatial-buffer range of 15 to 80 m (Fig. A2c). Correlations between unsanitary latrines and molecular targets per 100 mL of *mE. coli* and total Bacteroidales, however, were only marginally significant ($p \sim 0.05$) at distances ranging from 80 to 100 m. Sanitary latrine counts were correlated with Most Probable Numbers of *E. coli* within 40 to 70 m from a pond (Fig. A2d). Adenovirus copy numbers were not correlated with any type of latrine (total, unsanitary or sanitary) at any distance (Fig. A2b, c, and d).

The optimal distance for correlations between latrine counts and fecal bacteria was 60 m, since this was the only radius that resulted in significant correlations with all three total fecal indicator bacteria,

cultured *E. coli*, *mE. coli* and total Bacteroidales (Fig. A2b). The correlation between *E. coli* and the total number of latrines ($r_s = 0.40$) within 60 m of a pond was stronger than either unsanitary (0.33) or sanitary latrines (0.31) considered separately (Table 2). *mE. coli* and total Bacteroidales correlated to total latrine counts within 60 m of the pond with r_s equal to 0.37 and 0.36 respectively. Considering the number of unsanitary latrines only within the pond drainage basin resulted in correlations with fecal indicators *E. coli* ($r_s = 0.46$), *mE. coli* (0.37) and total Bacteroidales (0.41). These correlations were stronger than those using spatial-based latrine counts (Table 2), and contrast with the lack of correlation between fecal indicator bacteria and sanitary latrines counted within each pond drainage basin.

4. Discussion

4.1. Population and sanitation influences on fecal bacteria concentrations

There is an urgent need to better understand the role of the environment in the transmission of diarrheal disease in Bangladesh where an estimated 11% of all deaths are attributable to diarrheal disease (Streatfield et al., 2001). Exposure to pond water is a risk factor for diarrheal disease in rural Bangladesh (Emch et al., 2008; Ali et al., 2002), so understanding the sources of fecal contamination and environmental factors affecting the degree of pond fecal contamination is important in the development of a suitable strategy to limit human exposure to fecal waste. This study tested three hypotheses: 1) fecal bacteria in ponds originate primarily from humans rather than cattle; 2) concentrations of fecal bacteria and adenovirus are related to local population density; and 3) concentrations of fecal bacteria and adenovirus are related to the number of latrines around a pond, as well as the type of latrine. Both the frequency of detection and concentrations of human host-specific Bacteroidales exceeded the bovine host-specific Bacteroidales indicating that pond water fecal pollution in Char Para is overwhelmingly of human origin. The widespread presence of human adenovirus (41 out of 43 ponds) also supports this hypothesis. Although cattle are abundant in villages, cattle manure serves as fuel for cooking and is therefore collected and traded in rural Bangladesh, which may explain the relatively low, but

widespread abundance of bovine-specific markers. Fecal bacteria marker concentrations were positively correlated with population and latrine density within 45 m of the pond strengthening the quantitative link between human fecal sources and pond contamination. In contrast, copies per 100 mL of adenovirus were not correlated to population density, total latrines, or any other variables, suggesting that factors influencing adenovirus copy numbers are more complex than those influencing fecal bacteria concentrations.

Latrines were categorized as sanitary or unsanitary based on previously identified associations between diarrheal disease incidence and the number of unsanitary latrines surrounding a dwelling (Emch et al., 2008). In this study, the number of unsanitary latrines within each pond drainage basin was correlated to fecal bacteria concentrations ($p < 0.05$), whereas the number of sanitary latrines was not, demonstrating the negative impact of unsanitary latrines on neighboring surface water. This unsanitary latrine category included 22 (out of 99) make-shift above-ground latrines consisting of two bricks on the ground. The difficulty of finding these relatively secluded “latrines” suggests many more than were found may be present and causing pollution, highlighting the importance of properly functioning latrines to minimize fecal contamination of bathing ponds in densely populated villages similar to those in Bangladesh.

The number of unsanitary latrines within a pond drainage basin was more predictive of fecal indicator bacteria concentrations than spatial-buffer counts of any latrine type (total, sanitary, unsanitary) within a 60 m radius of each pond (Table 2). The fact that unsanitary latrine counts within pond drainage basins were more predictive is likely the result of topography surrounding each pond that determines the direction of latrine effluent and surface runoff flow. Latrines are often placed near ponds for esthetic reasons and these latrines will have the greatest influence on pond water quality.

4.2. Correlations between fecal bacteria and adenovirus markers

The cultured fecal indicator bacteria *E. coli*, and molecular markers for *mE. coli* and total Bacteroidales correlated strongly with each other indicating that any three of these assays may be used to evaluate levels of pond water fecal contamination. Total Bacteroidales molecular targets were consistently detected in higher numbers than *E. coli* genomes and thus may be advantageous in some settings for detecting lower concentrations of fecal contamination. Bacteroidales are obligate anaerobes and their persistence is sensitive to both oxygen and temperature (Bell et al., 2009). Walters and Field (2009) measured more rapid decreases in target copy numbers of human Bacteroidales than cultured *E. coli*, with T_{90} 's, the time required for a concentration decrease of 90% based on measured decay rates, of 1.9 and 7.9 days respectively (sewage-seeded, fresh water, and dark microcosms). In contrast to Bacteroidales, *E. coli* may be retained longer in the environment but may also re-grow with a recent study suggesting that some strains of *E. coli* may be endogenous to soil bacteria (Nautiyal et al., 2010). In the present study, the strong correlation between *E. coli* and Bacteroidales, however, indicates that these differential die-off and re-growth processes did not diminish the efficacy of Bacteroidales as a quantitative fecal indicator in surface waters. The ratio of culturable *E. coli* to *E. coli* genomes was not influenced by pond type and temperature, indicating that neither nutrient availability nor temperature affected the proportion of culturable genomes in the pond samples.

Although widely detected, adenovirus copy numbers were not correlated with any of the fecal indicator bacterial markers. Similarly, Bofill-Mas et al. (2010) found widespread occurrence of adenovirus in European fresh and marine waters (80 out of 132 samples positive), whereas only 3 of 15 laboratories found significant correlations between cultured *E. coli* cell numbers and adenovirus copy numbers ($p < 0.05$). These findings may be partially explained by the high stability of adenovirus in wastewaters with a measured T_{90} of 61 days

(Bofill-Mas et al., 2006). Walters et al. (2009) found that in contrast with human-specific Bacteroidales, enterovirus copy numbers were relatively stable and not susceptible to the effects of solar radiation. These differences in inactivation rates and processes between viruses and fecal indicator bacteria may explain the poor correlations typically observed in surface waters (Lee et al., 2004; Wilkes et al., 2009).

Significant positive correlations ($p < 0.05$) were observed between all fecal indicator bacteria and specific conductivity (Table 2). Conductivity is likely to rise with increased anthropogenic use of the pond catchment area, including salt inputs from human waste (Manahan, 2005) and the use of ash for washing (Hoque et al., 1995; Sengupta et al., 2008) coupled with pond water evaporation, and may therefore be a proxy for human usage and pond water contamination.

4.3. Implications for the spread of diarrheal disease

Most Probable Numbers of *E. coli* in the surveyed ponds exceeded the U.S. EPA recreational water quality limit (126 MPN/100 mL) up to 10,000 times (US EPA, 1986) and were in fact similar in concentration to fecal coliforms typically detected in untreated wastewater (Sinton et al., 1999; Passerat et al., 2011). A global epidemiological meta-analysis determined that for every 1 log₁₀ increase in *E. coli* the odds of acquiring diarrheal disease from recreational contact approximately doubles (Wade et al., 2003). Bathing and fishing ponds contained 1–2 log₁₀ less culturable *E. coli* than latrine ponds suggesting that diarrheal disease risk from exposure to latrine ponds is 2 to 4 times higher than for protected ponds. However, even fishing and bathing ponds had high levels of all fecal indicators, *E. coli*, *mE. coli*, total Bacteroidales and human Bacteroidales with median concentrations of 10³ (MPN), 10⁵, 10⁶ and 10⁵ (copies) per 100 mL respectively. These high levels of fecal pollution as well as the widespread detection of the human gastrointestinal pathogen adenovirus, and close proximity to human dwellings implicate these ponds as likely sources of persistent diarrheal disease. Human exposure to fecal contamination will therefore be limited by the continued upgrading and maintenance of latrines and the placement of these latrines outside of the immediate pond drainage basin.

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Appendix A. Supplementary data

Supplementary data to this article can be found online at doi:10.1016/j.scitotenv.2011.04.043.

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